Quantitative prediction of interactions in bipartite networks based on traits, abundances, and phylogeny

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Abstract

Ecological interactions link species in networks. Loss of species from, or introduction of
 new species into, an existing network may have substantial effects for interaction patterns.
 Predicting changes in interaction frequency while allowing for rewiring of existing inter actions, and hence estimating the consequences of community compositional changes is
 thus a central challenge for network ecology.

Interactions between species groups, such as pollinators and flowers or parasitoids and
hosts, are moderated by matching morphological traits or sensory clues, most of which are
unknown to us. If these traits are phylogenetically conserved, however, we can use phylogenetic distances to construct latent, surrogate traits and try to match those across groups,
in addition to observed traits. Understanding how important traits and trait-matching are,
relative to abundances and chance, is crucial to estimate the fundamental predictability of
network interactions.

Here we present a statistically sound approach ("tapnet") to fitting abundances, traits 14 and phylogeny to observed network data in order to predict interaction frequencies. We 15 thereby expand existing approaches to quantitative bipartite networks, which so far failed 16 to correctly represent the non-independence of network interactions. Furthermore, we use 17 simulations and cross-validation on independent data to evaluate the predictive power of 18 the fit. Our results show that tapnet is on a par with abundance-only, matching-centrality 19 and machine learning approaches. This approach also allows us to evaluate how well cur-20 rent concepts of trait matching work. Based on our results, we expect that interactions in 21 well-sampled networks can be well predicted if traits and abundances are the main driver 22 of interaction frequency. 23

24 Keywords: community, morphological trait, mutualist network, phylogeny, pollination

²⁵ Introduction

Network ecology has vastly increased our knowledge of ecosystems, delivered fascinating in-26 sights into their organization (e.g. the slow-and-fast-energy channels in Rooney et al. 2006), 27 and fostered speculation about co-evolution (e.g. Guimarães Jr et al., 2011). While arguably the 28 litmus test of any ecological understanding is a test of its predictive power (Houlahan et al., 29 2017), only recently have predictive models of network interactions started to emerge (e.g. Ives 30 and Godfray, 2006; Petchey et al., 2008; Crea et al., 2016; Rohr et al., 2016; Brousseau et al., 2018; 31 Pichler et al., 2020). After several decades of research on ecological networks we have learned 32 a lot about their patterns (e.g. along latitudinal or elevational gradients, across different types 33 of interactions) and their stability (see, e.g. Bascompte and Jordano, 2014; Moore et al., 2017), 34 yet so far we have had limited success in predicting interaction intensities or even aggregate 35 network structure (as pointed out by Vázquez et al. 2009; Olito and Fox 2015; Poisot et al. 2016; 36 Valdovinos et al. 2018, but see Pomeranz et al. 2019; Vizentin-Bugoni et al. 2020). 37

Some of the most successful attempts at predicting network structure were made in studies 38 of food webs, where body size ratios were found to be a good predictor of presence or absence 39 of predator-prey interactions (e.g. Allesina, 2011; Gravel et al., 2013; Pomeranz et al., 2019), and 40 allometric scaling of parameters allowed to fit a mechanistic model of optimal foraging theory 41 (the contingency model: Stephens and Krebs, 1986) to large networks (Beckerman et al., 2006; 42 Petchey et al., 2008). However, body size is not a suitable predictor for most other interaction 43 types, for example host-parasitoid interactions or interactions involving plants (Bascompte 44 and Jordano, 2014; Dormann and Blüthgen, 2017). Moreover, while network predictions should 45 ideally be based on theory, existing theoretical models such as the contingency model make 46 strongly simplifying assumptions (e.g. sequential encounter of resources, maximising of av-47 erage resource intake, ideal knowledge of resource availability, no within-guild interference: 48 Pyke, 1984; Stephens and Krebs, 1986), but still require large numbers of parameter values as 49 input, which are difficult to provide without a shortcut such as allometric scaling. Thus, we still 50 lack a general approach to predict various types of ecological networks. In addition, whereas 51 most existing models of ecological networks only predict binary network structure, models 52

that allow to estimate the intensity of interspecific interactions can glean more information
 from the data.

⁵⁵ While a theoretical model encompassing all types of interaction networks does not yet ⁵⁶ exist, there are obvious candidate factors for predicting interaction intensities, some at the level ⁵⁷ of the individual species (such as their abundance), some at the interplay of species, such as the ⁵⁸ match of interaction-relevant traits (Junker et al., 2010, 2013; Dehling et al., 2016). Since some ⁵⁹ of the relevant traits are difficult to measure, but show a phylogenetic signal, phylogeny can ⁶⁰ be used as a proxy for such unmeasured traits (Ives and Godfray, 2006; Pearse and Altermatt, ⁶¹ 2013; Morales-Castilla et al., 2015; Peralta, 2016).

There are several previous approaches to include traits, abundance and phylogeny into a 62 statistical method to analyze and possibly predict interactions. They differ in their scope (bi-63 nary vs quantitative networks), the type of network (bipartite vs. unipartite) and the statistical 64 method. For example, Ives and Godfray (2006), Pearse and Altermatt (2013), Rohr et al. (2016) 65 and Crea et al. (2016) all used regression models to predict binary networks based on phy-66 logeny and measured traits. Brousseau et al. (2018) improved on the model of Rohr et al. (2016) 67 by adding a larger number of traits as predictors and using a more flexible Generalized Addi-68 tive Model (GAM). Vázquez et al. (2009) and Olito and Fox (2015) chose a different statistical 69 approach in which they constructed matrices of interaction probabilites based on traits and 70 abundances and assumed that the observed interaction intensities are drawn from a multino-71 mial distribution with these probabilities. Most recently, Desjardins-Proulx et al. (2017) and 72 Pichler et al. (2020) modelled interaction networks using machine-learning algorithms, which 73 are highly flexible, but provide little information on the underlying mechanisms of the inter-74 actions. 75

While the above-mentioned approaches have had some success in describing ecological networks, they are faced with two issues, one statistical, the other ecological. First, most methods treat the interactions in a network as statistically independent (except Vázquez et al., 2009; Olito and Fox, 2015; Crea et al., 2016), although this assumption is likely to be violated. Any interaction with one species precludes the interaction with other species at the same time; thus, one more observation here inevitably means one less there. Furthermore, depending on the type of interaction and method of data collection, the same individual may be observed
multiple times, again violating the independence assumption. Finally, interactions of different consumer species may be non-independent due to intraguild competition for resources,
which can cause shifts in species' preferences (e.g. Loeuille and Loreau, 2005; Spiesman and
Gratton, 2016). Such non-independence must be accounted for in order not to yield biased and
overconfident model estimates.

An ecological issue is how to represent the role of species traits and their matching. Models 88 based on linear regression assume that all traits and trait combinations have linear effects on 89 interaction probabilities, while machine-learning algorithms do not provide any information 90 on the mechanisms connecting traits to interactions. In reality, interaction intensities may 91 depend on the matching of quantitative trait values in nonlinear and possibly asymmetric 92 ways. For instance, large billed birds can feed on small seeds, but small billed ones not on 93 large seeds (e.g. Muñoz et al., 2017). Ideally, models of ecological networks should allow to 94 explicitly incorporate such trait-matching mechanisms, to correctly represent the ecological 95 mechanisms and provide accurate predictions (but see Sebastián-González et al., 2016, for an 96 implicit approach). 97

In this paper we present a statistical approach to analyzing and predicting interaction in-98 tensity, based on observed and phylogeny-based latent traits and their matching, alongside 99 abundances of each species. Unlike previous analyses, which were mostly explorative, we-100 assess our model's performance on independent network data that were not used for fitting. 101 For more ecological realism, and in extension to previous approaches, we provide symmetric 102 and asymmetric trait-matching functions. In contrast to all previous approaches for traits and 103 phylogeny, we account for the non-independence of observations in the network, using the 104 multinomial probability approach of Vázquez et al. (2009) and Olito and Fox (2015). Finally, we 105 assess the quality of our approach at the level of the individual link, not only at the level of 106 network patterns, as the aim is to predict a specific interaction. 107

Our approach can be used for prediction of, say, introduction or loss of species from a community, which may lead to a "rewiring" of interactions, or for predicting the effects of changes in abundances. Moreover, it can be used to quantify the importance of observed trait pairs for such predictions, such as the morphological compatibility of a plant's flower corolla
and a pollinator's proboscis, or the beak size of a frugivorous bird and the size of a fruit (see
also Pichler et al., 2020).

We first outline the idea and details behind our approach, then demonstrate its validity with simulated data, and finally use three hummingbird-flower networks to make predictions across habitats. In that case study, we also use alternative published approaches to gauge their performances on real-world validation data.

Methods

We call our approach "tapnet", as it uses traits, abundance-activity and phylogeny to predict 119 network interactions. Our background in pollination ecology makes it natural for us to think 120 of bipartite interaction networks, where one group's members (e.g. pollinators) interact with 121 members of another group (e.g. plants), but not within each group. However, the approach can 122 be similarly applied to functional group- or individual-based networks and probably extended 123 to one-mode networks, but that is beyond the scope of our study. In the case of analysing 124 individuals, species' average traits would be replaced by individual trait values. Abundances 125 could either be removed completely or replaced by some measure of individual activity. 126

Before explaining our approach in detail, here is an outline (Fig. 1): We developed a model 127 that outputs expected interaction probabilities based on traits, abundances and phylogenies. 128 We can now compare the output with an observed network and optimize model parameters so 129 as to maximize fit. In this way we estimate several (largely) ecologically interpretable param-130 eters. For prediction, we can use the fitted model together with new abundances (including 131 previously unobserved species) and yield expected interaction probabilities. While the role of 132 traits is probably clear, phylogenies are used to construct so-called latent traits to be matched 133 across groups. As a side effect, new species entering a community can be positioned in the 134 phylogeny and this position then feeds through to the actual predictions, in addition to the 135 (optional) observed traits. 136

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Traits refer to species-specific characteristics that have a counterpart in a trait of the other

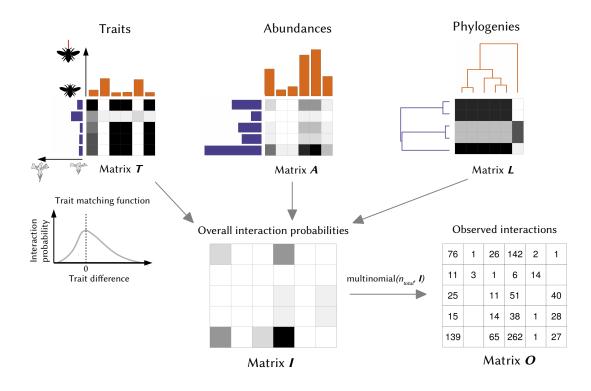


Figure 1: Conceptual overview of the tapnet approach. Tapnet combines information on trait matching (**T**), abundances (**A**) and phylogeny-derived matching latent traits (**L**) (top row) into a single matrix **I** of predicted interaction probabilities. The observed interaction network **O** is assumed to be drawn from a multinomial distribution with probabilities given by **I** and total number of interactions n_{total} equaling the observed number of interactions. Traits have to be provided in pairs across the two levels, which are then compared using a possibly asymmetric trait-matching function. Each input yields an independent matrix of interaction probabilities, which are then multiplied and re-scaled to yield the final interaction matrix **I**. Matrix cells are filled in different shades of gray to indicate interaction probabilities from zero (white) to one (black). Some values which are close to zero appear white.

group, and the analyst must provide these traits in matching format (i.e. corolla and proboscis
length, or phenologies during a season, quantified in such a way that the same value for both
would imply a perfect match). The model can be employed without trait information.

Abundance-activity (henceforth abundances) can be any measure proportional to the probability of encountering a species in the field. For instance, a plant with an attractive scent would be more conspicuous to a pollinator than a non-odorant one. And an abundant bee would be encountered more often than a rare one, if they move at the same rate (hence abundanceactivity). If the trait-pairs responsible for interaction activity are known and provided to tapnet, the role of actual abundances can be quantified, otherwise the two will remain confounded (as in any other analysis).

Phylogenies for each group are used as building blocks for unobserved (latent) traits moderating species interactions. They add complementary information to the observed traits (Pearse et al., 2013), and may be able to suggest possible additional traits that underlie the observed interactions.

We present the approach starting with the statistical goal and then going into ever more detail. We provide R-code for simulating "tapnet" data, for fitting observed networks, for assessing model fit and for predicting to new abundances and traits in the package tapnet.¹

155 Likelihood

¹⁵⁶ Our approach is to fit a (non-standard) statistical model to observed data on interaction net-¹⁵⁷ works. While we can use several networks simultaneously to fit the model, we restrict our ¹⁵⁸ outline here to the simplest case of only one observed network. Let us call the observed inter-¹⁵⁹ action matrix **O** of dimension $m \times n$ (*m* rows by *n* columns).

We compare **O** with our model predictions **P** by means of a multinomial distribution (eqn 1), as entries in that interaction matrix are non-independent. (Entries in **O** are integers, typically number of observed interactions per standardized observation effort. For continuouslyvalued observations, a Dirichlet distribution could probably be employed: Crea et al. 2016.) To do so, we represent **O** as a vector of length *mn*, rather than a matrix. Formally, we predict the

¹https://github.com/biometry/tapnet

entries in **O**, o_i , based on the total number of observed interactions, $n_{\text{total}} = \sum_{i=1}^{mn} o_i$, and the vector of predicted interaction probabilities, \hat{p}_i , which are the output of our model:

$$f(o_1, \dots o_{mn}; n_{\text{total}}, \hat{p}_1, \dots, \hat{p}_{mn}) = \frac{n_{\text{total}}!}{\prod_i o_i!} \prod_{i=1}^{mn} \hat{p}_i^{o_i}$$
(1)

¹⁶⁷ Model components

Our model yields a matrix of predicted interactions, $\mathbf{P} = (\hat{p}_{i,j}) \in \mathbb{R}^{m \times n}$, as (re-scaled) Hadamard (= element-wise) product of three prediction components: abundance-based expected probabilities **A**, trait-matching based expected probabilities **T**, and latent-trait-based expected probabilities **L**, each scaled to sum to 1:

$$\mathbf{P} = \mathbf{A} \circ \frac{(\mathbf{T} \circ \mathbf{L})^{\delta}}{\sum (\mathbf{T} \circ \mathbf{L})^{\delta}}.$$
 (2)

A is the matrix of abundance-based interaction probabilities based on the cross-product of 172 normalized species abundances vectors for the lower o_A^l and higher o_A^h trophic level (note that 173 we reserve capital letters for matrices): $\mathbf{A} = \frac{o_A^l}{\Sigma o_A^l} \times \frac{o_A^h}{\Sigma o_A^h}$. Thus, \mathbf{A} represents the probability 174 of an interaction for each cell based only on the relative abundances of the different species: 175 we would expect more interactions among common species than among rare species. $o_A^{l,h}$ is 176 based on independently measured abundances, not simply the marginal totals of the observed 177 matrix **O**. The free exponent, $\delta \in (0, 1)$, allows the optimization to give more or less weight to 178 traits relative to abundance, and thereby also serves as a quantification of the importance of 179 abundance within the fit. As a side-effect, the abundance-only predictions are a limiting case 180 of tapnet, for $\delta = 0$. The denominator summing over $(\mathbf{T} \circ \mathbf{L})^{\delta}$ is necessary to re-normalize this 181 term to sum to 1, on a par with A. 182

T is the matrix of expected interaction probabilities based on the degree to which observed traits match between species of the different groups. If, say, the proboscis of a pollinator is a bit too short or too long compared to the depth of the corolla of a plant species, then interactions become less likely than perfectly matching lengths. We define a (single parameter) trait-matching function for a pair (i, j) of trait values $f_t(t_i^l, t_j^h)$ by the Gaussian function:

$$f_t(t_i^l, t_j^h, \sigma) = \frac{1}{\sigma\sqrt{2\pi}} e^{-\frac{(t_j^h - t_i^l)^2}{2\sigma^2}}.$$
(3)

Alternatively, we can define an asymmetric matching function (a log-normal with its mode shifted to 0), as too long a proboscis is no obstacle to an interaction, while one too short is:

$$f_t(t_i^l, t_j^h, \mu) = \frac{1}{(t_j^h - t_i^l + e^{\mu - 1})\sqrt{2\pi}} e^{-\frac{1}{2}(\ln(t_j^h - t_i^l + e^{\mu - 1}) - \mu)^2},$$
(4)

¹⁹⁰ where $e^{\mu-1}$ is a parameter determining the shape of the function, in a similar way to σ in eqn 3. ¹⁹¹ It is fit during the model optimization (see further below). Other functions could of course ¹⁹² be used, especially when the relationship between trait values and interaction probability is ¹⁹³ known. The trait-matching functions is computed for all species, i.e. the inputs are in fact ¹⁹⁴ vectors of trait values for the lower (\mathbf{t}^l) and higher level (\mathbf{t}^h). The elements of \mathbf{T} , (τ_{ij}), are the ¹⁹⁵ f_t -values computed using eqns. 3 or 4: (τ_{ij}) = $f_t(t_i^l, t_j^h)$.

Latent-trait-based interaction probability matrix **L**, finally, is the most complex matrix of expected interaction probabilities. Its function is primarily to improve predictions, as abundances and observed traits are unlikely to be sufficient to capture the information present in the observed interaction matrix. Just like **T**, also **L** is found by matching traits of one level to those of the other by optimizing the σ -parameter of its Gaussian trait-matching function (eqn 3). In this case, however, these traits are unobserved (latent) and are constructed as part of the optimization process from phylogenetic eigenvectors, as explained in the next section.

203 Constructing matching latent trait values

The reasons why a species from one group interacts with one from another may be extremely complex. Bees, for example, may rely on visual and olfactorial signals, as well as innate preferences and learned behavior, for selecting a flower to visit (Chittka and Raine, 2006). On the other hand, plants may increase nectar production in response to sensing pollinator's wingbeats (Veits et al., 2019). If these traits can be measured, they can be used in tapnet as observed traits. In many cases, however, we may have little chance to identify, let alone measure, the traits involved. In such cases, one can "invent" trait values with the sole aim of improving the
fit of the model.

Computationally it is more efficient, and ecologically more satisfactory, if there is a basis 212 for computing such latent trait values (similar to geographic distances in spatial models used 213 in joint species distribution models: Warton et al., 2015). In our case, we assume latent traits 214 to summarize traits correlated with phylogeny. These latent traits typically exhibit no phy-215 logenetic signal anymore, and the phylogenies are only used as an efficient way to generate 216 orthogonal vectors. This approach has several benefits. Firstly, it allows us to introduce new 217 species when predicting from the fitted tapnet model, as we can compute their phylogenetic 218 position relative to the other species in the group, and hence also the value of their latent 219 trait. Secondly, the resulting latent trait may, upon mapping it to the phylogeny, suggest a 220 hypothesis about an actual trait behind it. 221

Technically, constructing a latent trait within a group is straightforward if a phylogeny 222 is available. Alternatively, a taxonomy can be used (Clarke and Warwick, 1999). From such a 223 phylogeny one can compute phylogenetic eigenvectors (Guénard et al., 2013), i.e. for k species 224 k-1 vectors that are orthogonal to each other and represent the information of the phylogeny 225 (similar to the way a Principal Component Analysis summarizes the information in a data set): 226 the first phylogenetic eigenvector accounts for the largest genetic differences in the tree, the 227 second for the largest in the remaining variation, and so forth. Mathematically, a phylogenetic 228 tree can be represented as a (cophenetic) distance matrix, and an eigenvalue decomposition of 229 said distance matrix yields the phylogenetic eigenvectors. 230

For each group, we can now define a latent trait vector **l** as a linear combination of the phylogenetic eigenvectors $\mathbf{p}_i, \dots, \mathbf{p}_{k-1}$, one for each trophic level:

$$\mathbf{l}^{l} = a_{1}\mathbf{p}_{1}^{l} + a_{2}\mathbf{p}_{2}^{l} + a_{3}\mathbf{p}_{3}^{l} + \dots + a_{m-1}\mathbf{p}_{m-1}^{l}$$
$$\mathbf{l}^{h} = b_{0} + b_{1}\mathbf{p}_{1}^{h} + b_{2}\mathbf{p}_{2}^{h} + b_{3}\mathbf{p}_{3}^{h} + \dots + b_{n-1}\mathbf{p}_{n-1}^{h}$$
(5)

For computational reasons, we will only use the first few (3-5) phylogenetic eigenvectors. When \mathbf{l}^{l} and \mathbf{l}^{h} match, the probability of an interaction is high, just like for observed traits ²³⁵ \mathbf{t}^{l} and \mathbf{t}^{h} above. The values of a_{i} and b_{i} have to be found by optimisation. Since the values of ²³⁶ \mathbf{l} are void of meaning, one cannot assume that the latent vectors align optimally. Therefore, ²³⁷ a shift parameter is fitted for the higher trophic level (b_{0}). Analogous to \mathbf{T} , the elements of ²³⁸ \mathbf{L} , (ℓ_{ij}), are the f_{t} -values computed using eqn 3 based on the latent traits: (ℓ_{ij}) = $f_{t}(l_{i}^{l}, l_{j}^{h})$.

Fitting the model

The tapnet model outlined above can now be fitted to the observed interaction network by adjusting several parameters. As input the model requires the paired observed traits (0 to many); the phylogeny of each group; the abundance vector for each group, and for computation of the likelihood, the observed interaction matrix. The model parameters are

1. the width of the trait matching function (σ in eqn 3) for each pair of traits;

245 2. the width of the trait matching function for the latent traits; and,

²⁴⁶ 3. two vectors of parameters for the construction of the latent trait (eqn 5).

In the optimization using the standard Nelder-Mead algorithm, we used a few tricks to increase the reliability of the model. To ensure identifiability, we constrained a_1 to be positive (i.e. defined it as e^{a_1}). Otherwise the exact same values with inverted signs would yield the same fit. Furthermore, any multiple of \mathbf{a}^l , \mathbf{a}^h would yield the same prediction. Therefore we standardized both \mathbf{l}^l and \mathbf{l}^h before entering them into the trait-matching function (eqn 3).

²⁵² When fitting multiple networks simultaneously or when using different networks for op-²⁵³ timization and prediction, we run into the problem of having different species present in each ²⁵⁴ network. In such cases we first calculate the eigenvectors of the phylogenetic tree containing ²⁵⁵ all species from all networks. In a second step, we select from these eigenvectors those that ²⁵⁶ are most relevant for the respective network, i.e. the eigenvectors most closely correlated with ²⁵⁷ each of the eigenvectors of the tree containing only the species of this particular network. We ²⁵⁸ then fit parameters only for the selected relevant eigenvectors.

Assessing model quality

Optimizing the model parameters immediately yields the model's likelihood. Additionally, we 260 may be interested in the latent variables, as they code the (combination of) trait(s) missing 261 in our observed data. For simulated data, we can compare the reconstructed latent traits with 262 those actually simulated. Predicted and observed interactions were additionally compared us-263 ing other distance measures, such as the Pearson correlation or Bray-Curtis distance, or by 264 summarizing the network structure by indices (such as nestedness or specialization). For net-265 work indices, we drew 1000 realizations from the fitted multinomial distribution of each sim-266 ulated network (see below) and computed network indices for these. Then, we computed on 267 which quantile of these 1000 realization the observed network's index lay (sometimes called 268 the 'posterior p-value': Gelman, 2005). Ideally, this value should be 0.5, indicating no bias in 269 indices in the fitted network. 270

271 Simulations

To assess how performance of the tapnet model varies with the characteristics of the data 272 used for fitting, we performed two simulation experiments. The first aimed to evaluate the 273 model's goodness of fit to the interaction network used for estimating parameters, while the 274 second was designed to test the model's accuracy of prediction to a new network. In both 275 experiments we varied six parameters (Table 1) using Latin hypercube sampling (McKay et al., 276 1979). Ranges of numerical parameters were divided into 500 equally spaced intervals, and a 277 single random value was drawn from each interval. For parameters with integer values (e.g. 278 number of observed traits), drawn values were rounded to the nearest integer. In the case 279 of categorical parameters, we randomly sampled 500 times with replacement from the set of 280 possible values. (A pre-run with only 100 samples yielded virtually identical results, indicating 281 that 500 runs are sufficient.) Since the number of observed traits varied between zero and 282 four, we drew four sets of 500 values for the width parameter of the trait matching function. 283 Depending on the number of observed traits of the respective parameter combination, we used 284 only a subset of the four values to calculate matrix \mathbf{T} (or none at all with zero traits). 285

Table 1: Parameters varied in the simulation experiments used to assess the model's goodness of fit. The "width" parameter defines the sensitivity of the probability of an interaction to mismatches between traits; small values demand very neat matching for an interaction to be likely.

Parameter	Туре	Range or possible values
Total number of observed interactions n_{total}	integer	50 - 1000
Number of observed traits	integer	0 - 4
Type of trait-matching function for observed	categorical	normal or shifted log-normal
traits		
Shape of abundance distribution	categorical	uniform or log-normal
Width parameter of trait matching function for	continuous	0.05 - 1
observed traits		
Width parameter of trait matching function for	continuous	0.05 - 2
latent traits		

For each of the 500 parameter combinations, we simulated a data set consisting of relative abundances, phylogenetic trees and pairs of matching traits. Species abundances were either all set to the same value (1/m or 1/n, respectively) or drawn from a log-normal distribution with parameters $\mu = 0$ and $\sigma^2 = 1$ and standardized to a sum of one. Trait values were likewise drawn from a log-normal distribution with $\mu = 0$ and $\sigma^2 = 1$. We simulated phylogenetic trees using the function "pbtree" from R package "phytools" (Revell, 2012) with a speciation rate of 1 and extinction rate of 0.

For both experiments, we simulated phylogenies and traits of 30 and 60 lower- and higher-293 trophic level species, respectively, for each parameter combination. These data were used to 294 construct a matrix of interaction probabilities P according to the "tapnet" model as described 295 above. While the width parameters of the trait matching functions for observed and latent 296 traits were systematically varied between simulations, the latent trait linear combination pa-297 rameters a_i were set to a value of 1 in all simulations. From the interaction probabilities we 298 constructed a network of simulated interactions with total number of interactions n_{total} by 299 drawing from a multinomial distribution with probabilities $\hat{p}_{i,j}$. 300

For the first experiment, we then randomly selected 15 (lower trophic level) and 30 (higher trophic level) species. To these data, we fit tapnet and assess goodness of fit.

For the second experiment, we randomly drew *twice* a random set of 15×30 species. Again, tapnet was fit to the first, but then predicted to the second set. On average, these two sets share half of their species and 25% of their interactions. It thus represents a strong test of
 predicting to new data. As measures of goodness of fit, we calculated the Bray-Curtis similarity
 and Spearman rank-correlation between the entries of simulated and predicted networks.

³⁰⁸ Independently observed vs network-derived abundances

The majority of published interactions networks does not provide independent estimates of 309 the abundances of each species (o_A^l and o_A^h for lower and higher trophic level, respectively; 310 see eqn 2). In current network analyses, it is thus customary to use network-derived, marginal 311 totals of the network matrix (O_i . and O_{i} , respectively) as plugin instead (e.g. to formulate null 312 model expectations: Vázquez and Aizen, 2003; Barber, 2007; Blüthgen et al., 2007; Dormann 313 et al., 2009). However, these marginal totals carry the imprint of network structure. In one ex-314 treme, pollinators in the region may simply not be attracted by the flowers in the patch under 315 consideration and hence are not present in the network. Or, in the case of antagonistic net-316 works, a parasite may reduce the population size of its host to such a degree that interactions 317 are hardly observed, although their intensity is very high (e.g. Barbosa et al., 2017). 318

We investigated the consequences of using independent vs network-derived abundances on prediction quality with the simulated data. For the same simulated data, we once fit tapnet with the simulated independent abundances and once with the marginal totals of the simulated interaction network. We then predict to either the independent abundances of the second simulated network vs its marginal totals. We expect that network-derived abundances will lead to a better model prediction, simply because they contain information on the structure of the test network.

326 Case study

As a demonstration, we use the case study of Tinoco et al. (2017), who compiled data on hummingbird-pollination networks in three different habitats (forest, shrubland and a cattle farm) in the southern Ecuadorian Andes. The data published alongside the paper² include traits for both plants and hummingbirds, as well as external abundance data. These networks

²http://dx.doi.org/10.5061/dryad.j860

are unusually intensively sampled, with 1288, 3979 and 2405 interactions in each of the three
 habitats, respectively, across 32 plant and 14 hummingbird species, some occurring only in one
 habitat.

In the case study, we additionally compare the predictive performance of tapnet with three different alternative approaches: abundance-only, trait-matching-and-phylogeny GAM following the ideas of Brousseau et al. (2018), and a similar model using random forest (see supplementary material for R-code and detailed results).

The abundance-only model can be seen as a baseline: it uses only the information on the activity/abundance of the *m* lower level and *n* higher level species in the validation data *v*, specifying the estimated probability of interactions: $\widehat{\mathbf{P}}_{v} = \frac{\mathbf{o}_{v}^{l}}{\sum_{i=1}^{m} o_{v,i}^{l}} \times \frac{\mathbf{o}_{v}^{h}}{\sum_{j=1}^{n} o_{v,j}^{h}}$. Multiplying this with the number of observed interactions, $N_{\text{total}} = \sum o^{l} = \sum o^{h}$, yields the predicted interaction intensity. Only improving on this model demonstrates explanatory power of traits and their matching.

Following the approach of Brousseau et al. (2018), we fitted a negative-binomial GAM us-344 ing 2-D-splines on the first phylogenetic eigenvectors of each group, the same for the second 345 eigenvectors, the observed trait values per species of each group and the squared difference 346 between traits (representing trait matching). While the original approach used traits and phy-347 logenies to predict binary networks, here we predicted quantitative interaction matrices and 348 additionally used the abundances as predictors. Spline complexity was set to k = 3 for uni-349 variate and k = 20 for 2D-splines, and an additional shrinkage was imposed by setting gamma 350 to 1.4, both following Brousseau et al. (2018). Note that this approach, as well as the next, 351 implicitly assumes entries of the interaction matrix to be (conditionally) independent (see dis-352 cussion). 353

The random forest approach was run using default setting (i.e. 500 trees, trying the rounded down square root of number of predictors at each split). It was provided with the same information as the previous GAM but using all phylogenetic eigenvectors; it serves as a comparison of the algorithm's flexibility, as it allows for interactions among the predictors.

³⁵⁸ While tapnet can fit several networks simultaneously (see supplementary material), we ³⁵⁹ employed it akin to the other approaches in a cross-validation setting: the models were fitted

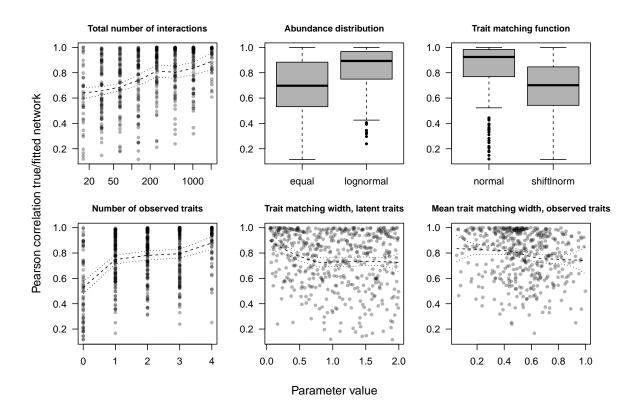


Figure 2: Pearson correlation between fitted model prediction and the simulated interaction network. Scatter plots show the correlation coefficient ρ as a function of six model parameters whose values were varied simultaneously using Latin Hypercube sampling. Data were simulated for 500 parameter combinations. Networks were of size 15 × 30 species. Lines represent local weighted smoothers and their 95% confidence interval to indicate trends in the simulations.

³⁶⁰ to one network, and then predicted to the two others in turn. Results were compared using

³⁶¹ the correlation between predicted and observed interactions.

Models were fit in R using packages mgcv (Wood, 2006) and ranger (Wright and Ziegler,

³⁶³ 2017); see supplement for R-code of simulations and case study.

364 **Results**

365 Simulations

In the first simulation experiment, the correlation between observed and fit networks was overall only moderate (mean value r = 0.76 across all 500 parameter combinations). A strong effect of the number of observations was detectable, yielding high correlations between ob-

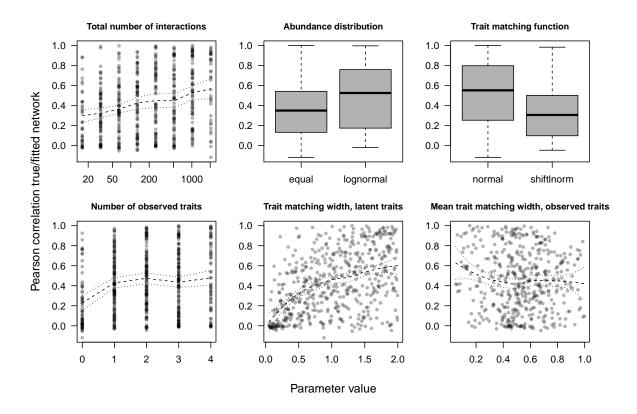


Figure 3: Pearson correlation between model predictions and a new interaction network simulated with identical parameter values. Scatterplots show the correlation coefficient ρ as a function of six model parameters whose values were varied simultaneously using Latin Hypercube sampling. Data were simulated for 500 parameter combinations. Lines represent local weighted smoothers and their 95% confidence interval to indicate trends in the simulations.

served and fit networks ($\bar{r} > 0.83$) for networks with more than 0.5 (i.e. 250 interactions in a 369 network with $15 \cdot 30 = 450$ cells) observations per number of cells (Fig. 2, top left). Correlation 370 coefficients were lower for networks with equal than for log-normal abundances of all species, 371 and higher for the 'normal' trait-matching functions than shifted log-normal. With increasing 372 number of observed traits, correlation coefficients also increased, indicating the usefulness 373 of observed traits for thinly sampled networks. The two trait-matching function parameters 374 (trait matching width of latent and of observed traits) did not seem to have a clear effect on 375 the correlation between observed and predicted networks. Patterns of variation in Bray-Curtis 376 similarity of observed and predicted networks were similar to those for Pearson's r (results not 377 shown). 378

In the second simulation experiment, with prediction to a new network, patterns were very similar to those in the fitting-evaluation, with overall lower correlations ($\bar{r} = 0.42$, and

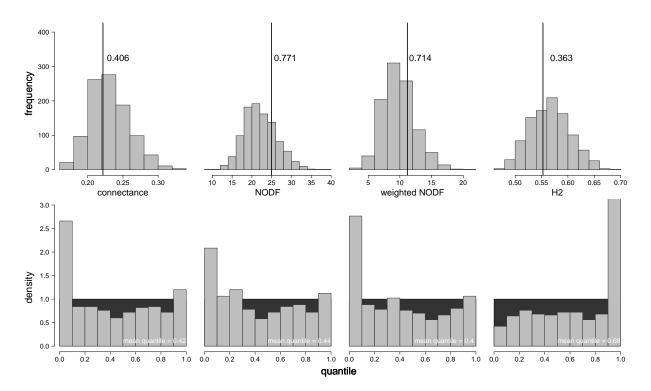


Figure 4: *Top row:* Example of network indices for 1000 realizations of one fitted network, compared with the simulated true value (vertical line; number gives quantile of true value within this frequency distribution). *Bottom row:* Quantiles of true network index values across the 500 simulations of experiment 1. Black background indicates ideal, uniform distribution. Despite a good overall match, spikes indicate a consistent underestimation of specialisation and hence lower connectance, lower nestedness and higher specialisation in the observed network.

for densely sampled networks $\bar{r} = 0.49$). Correlation coefficients depended strongly on the number of observations, abundance distribution and type of trait-matching function employed (Fig. 3). All other parameters had much less effect, and the difference between the two traitmatching functions can thus be seen as two almost separate sets of points in all plots apart from the categorical abundances.

³⁸⁶ Network indices for random realizations of the fitted network were very similar to those of ³⁸⁷ the observed, usually falling within the 95% confidence interval of the null model (Fig. 4). To ³⁸⁸ assess the coverage of the fits and their potential bias, we computed posterior *p*-values for each ³⁸⁹ index as mean quantile (see methods: Assessing model quality). Across the 500 simulations of ³⁹⁰ the first experiment, mean posterior *p*-values were: connectance 0.42; NODF 0.44; weighted ³⁹¹ NODF 0.40; and H'₂ 0.68, all indicating a slightly too generalist estimation of network structure ³⁹² (Fig. 4, bottom row). Table 2: Pearson correlation coefficients between fitted and observed hummingbird-flower network (Tinoco et al., 2017) for the four approaches, and their mean across the three habitats. Approaches are sorted by ranking in their cross-validation performance (Table 3).

Approach	Forest	Shrub	Farm	Mean
abundance-only	0.25	0.09	0.49	0.28
random forest	0.93	0.93	0.91	0.92
tapnet	0.57	0.60	0.65	0.61
GAM	0.56	0.29	0.40	0.42

³⁹³ Difference between independent and network-derived abundances

The fit of tapnet to the simulated data was slightly improved by using the marginal totals, 394 rather than the independently "observed" abundances ($\bar{r} = 0.76 \pm 0.22$ (1 sd) for independent 395 abundances compared to $\bar{r} = 0.83 \pm 0.15$ with marginal totals). Also the prediction to the 396 second simulated network, where half of the species were previously unobserved, improved 397 markedly from $\bar{r} = 0.41 \pm 0.30$ for independent abundances to $\bar{r} = 0.62 \pm 0.25$ with marginal 398 totals of the new network. Note that using only the marginal totals of the test network (with-399 out traits or phylogeny) to predict interactions already fitted and predicted well, with a high 400 correlation between marginals and the new network interactions ($\bar{r} = 0.66 \pm 0.17$). 401

402 Case study

The tapnet model fitted the three hummingbird-flower networks better than the abundance-403 only model and the GAM, but worse than the random forest approach (Table 2). On cross-404 validation, when fitting to one and predicting to the other two networks in turn, the tapnet 405 approach was no better or worse than the abundance-only and the random forest approach, 406 but substantially better than the GAM (Table 3). Interestingly, all approaches except the trait-407 neutral, abundance-only lost dramatically in performance from train to test data. The GAM, 408 for example, fitted the data moderately, but held no predictive power for the test data. Ran-409 dom forest, although reporting an extremely good fit, decreased to the level of tapnet and 410 abundance-only on the test data. This drop in performance from fitting to predicting suggests 411 that all statistical approaches overfitted, sometimes heavily. 412

Table 3: Pearson correlation of cross-validating by predicting with a model fitted to one habitat on the other habitats (indicated by \rightarrow). F, S and C are for forest, shrubland and cattle farm, respectively. For cross-validation log-likelihoods, which show the same result, see supplementary material.

Approach	$F \rightarrow S$	$F \rightarrow C$	$S {\rightarrow} F$	$S{\rightarrow}C$	$C{\rightarrow}F$	$C \rightarrow S$	Mean
abundance-only	0.09	0.49	0.25	0.49	0.25	0.09	0.28
random forest	0.23	0.36	0.33	0.14	0.40	0.17	0.27
tapnet	0.21	0.46	0.12	0.53	0.15	0.11	0.26
GAM	-0.01	0.33	0.26	0.13	-0.01	-0.02	0.11

Discussion

Over the last years, two main, not mutually exclusive lines of modeling approaches to predict 414 network structure have emerged (as reviewed in Valdovinos, 2019). Neutral models assume all 415 species to be similar and generalist, and hence describe an expectation for network structure 416 primarily based on sampling intensity and abundance distributions. Indeed, such approaches 417 are often used as null model against which to gauge the effect of interaction preferences (e.g. 418 Vázquez and Aizen, 2003; Blüthgen et al., 2006; Dormann et al., 2009). In contrast, interaction 419 constraint models focus on why some links are not present, typically using species traits and 420 their match across the groups as explanatory features (Santamaría and Rodríguez-Gironés, 421 2007; Bartomeus et al., 2016). In the approach presented here, we combine both by allowing 422 abundances to contribute or even entirely dominate the prediction, but use matching between 423 observed traits and between latent traits as constraints. 424

The results so far are both promising and sobering. On the one hand we demonstrated that 425 we can fit the observed data well with the tapnet approach, and predict reasonably in sim-426 ulated data. We can also predict with moderate accuracy networks from other habitats, and 427 that the tapnet approach did that better than some previous attempts to combine observed 428 and latent traits (Brousseau et al., 2018). On the other hand, by far the most important pre-429 dictor for our case study was species abundance, which makes it easy for any approach and 430 prevents the trait-matching strength of tapnet to play out. Thus, a neutral model, using only 431 abundances, was as good in prediction (but not fit), and at the same time much simpler, than 432 tapnet or random forest. We think that at least two factors contribute to this finding: (1) the 433 hummingbird-flower-network is not very specialized, despite featuring the most spectacular 434

sword-billed hummingbird (*Ensifera ensifera*), whose pollination-adapted bill is longer than its body, and (2) when abundance is very important, the log-likelihood becomes very shallow and difficult to improve upon. Also, abundance may be correlated with a trait relevant for interactions, although in this case there was no correlation between bill/seed size and abundances (r = 0.019 and -0.15, respectively; see supplement).

In simulations with many (> 500) observations, tapnet predictions were very reasonable, 440 while in the case study all approaches fared relatively poorly. Our simulations included pro-441 cesses deemed to be most important for determining network structure (compared to the list 442 in Valdovinos, 2019). This suggests that either mutualist networks may simply be extremely 443 noisy and under low evolutionary pressure, or that the current trait-matching concepts are 444 not good enough for describing, across networks, the processes that drive interactions. This 445 suggests a strong context-dependence of interactions, depending probably substantially on in-446 traspecific trait variation (Laughlin et al., 2012), behavioral complexity (Kaiser-Bunbury et al., 447 2010; Morán-López et al., 2020), competition within guilds (Vandermeer, 2004; Saavedra et al., 448 2013), non-linear frequency-dependence (Benadi and Pauw, 2018) and environmental condi-449 tions more generally (see Valdovinos, 2019, for review). Future research across many different 450 networks has to show whether abundance is consistently such an important predictor for in-451 teraction frequencies. 452

⁴⁵³ The chicken-and-egg problem of abundances and network structure

Not all studies record independent abundances, e.g. by estimating floral cover, sweep-netting 454 insects or alike, and this is more common for birds than for insects. Without such independent 455 abundance data, its role for determining interaction frequencies cannot be determined. One 456 frequent "solution" is to use the observed interactions of each species (the marginal totals of 457 the interactions matrix) as surrogate for its abundance. This approach has been rightly criti-458 cized as confounding the effect of abundances on network interactions with the effect of net-459 work structure on abundances, i.e. the chicken-and-egg problem of network interactions (Fort 460 et al., 2016; Dormann et al., 2017). This conflict was also detectable in our simulations, where 461 the tapnet-prediction to the test data was substantially improved by using the test network's 462

marginal total as predictor. Clearly, these surrogate abundances carry some information, be yond abundance, on network structure and hence interaction intensity.

Flower-visitation networks are, in general, only moderately "ecologically specialized" (sensu 465 Armbruster, 2017) (see, e.g., Blüthgen et al., 2007; Schleuning et al., 2012; Zanata et al., 2017), 466 suggesting that neither plants nor pollinators depend crucially on a specific (set of) species to 467 interact with. As a consequence, network structure and species abundances are strongly linked. 468 In these cases, marginal totals may arguably be used in lieu of independent abundances. Our 469 case study on a plant-pollinator system shows that this is not the correct approach. Indeed, 470 when predicting the fitted model once with the external and once with the marginal abun-471 dances, we find a dramatically better performance for the marginal abundances (across the 472 three habitats, predictive correlation is $\bar{r} = 0.75$, compared to a meager $\bar{r} = 0.26$ with inde-473 pendent abundances: see supplementary material). Clearly, marginal abundances contain an a 474 priori unknown amount of information, representing the outcome of within-guild interactions, 475 variable activity of individuals, selection cues (scent, visual signals) and so forth. 476

477 How independent are observed interactions?

We believe that treating observed interactions as independent data is statistically incorrect (see introduction), thus questioning the likelihood used by Rohr et al. (2016) and Brousseau et al. (2018), as well as the implicit independence assumption in Pichler et al. (2020).

Therefore, we here used a (network-wide) multinomial distribution, as had been suggested 481 by Vázquez et al. (2009), accommodating the compositional nature of the data. While it rec-482 ognizes the non-independence of observations, it does not thereby automatically capture the 483 processes behind it. That means, while the inference based on this distribution is probably 484 correct, our multinomial approach may not result in better predictions until the drivers of 485 non-independence are represented in the model (e.g. intra-guild competition). Additionally 486 this approach has two statistically relevant implications. First, we only get a single likelihood 487 value for a network, rather than *nm*. As a consequence, secondly, we assess the fit of the entire 488 model, without any chance of adapting only the fits of some specific species or interaction, 489 as is the case for the GAM and random forest approaches used here for comparison. While 490

we regard it as a more correct representation of the data, it also severely limits the type of
 statistical approaches that can be used for predicting network interactions.

Olito and Fox (2015) focus on a comparison of network indices produced by their predictive approach with those of the observed network. They conclude that even similar networks may have rather different index values, and different networks similar indices, making such indices a poor target for optimization. In our simulated networks, we were able to fit networks so that their index values centered on the observed value. This illustrates that tapnet did manage to fit network indices in line with the observations as a by-product of the multinomial likelihood *in principle*.

⁵⁰⁰ Traits, observed and latent

Following the lead of previous studies (in particular Rohr et al., 2016), tapnet uses phylogenetic 501 information to construct (not necessarily phylogentically conserved) latent traits to improve 502 fit to data. Clearly these are only a statistical placeholder for actual but unobserved ecologi-503 cal traits. Beyond the obvious but difficult to measure sensory interaction cues (Junker et al., 504 2013), also traits related to optimal foraging should be considered here - from both groups 505 of interacting species. As Pyke (2016) exemplifies, the fitness benefit of pollination for plants 506 depends on pollen-transfer efficiency of pollinators, and too high nectar rewards may give an 507 incentive to inefficient visitors. Latent traits may thus reflect a complex and fine-tuned pair of 508 matching sets of traits, without obvious interpretation. 509

A corollary of the possibility of fitting trait-pairs is also risk of identifying spurious char-510 acteristics of species (Mlambo, 2014). At present, interaction traits are almost exclusively mor-511 phological (see Junker et al., 2013, for an exception), but phenology can be incorporated simi-512 larly. In the future, both ecophysiological and genetic sampling may become sensitive enough 513 to extend research into interaction traits for example related to vision in the ultraviolet (e.g. 514 Rae and Vamosi, 2013), scent (e.g. Wright and Schiestl, 2009), or ultrasonic sound (e.g. Simon 515 et al., 2019). Until such devices become available, latent traits are a statistical stand-in for what 516 really makes species interact. 517

518 Conclusion

The approach we presented here predicts network interactions for new networks, conditional on trait, abundance/activity and phylogenetic data for the new network. It is flexible enough to include any type of function translating trait-matches into interaction probabilities. As a side-effect, it quantifies the importance of abundance relative to traits for network interactions. Future applications will have to assess the importance of traits across different types of networks, testing the assumption of many network studies that traits are the driving force of network interactions.

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Author contributions

⁵³³ Conceptualisation: GB, CFD, JF, DPV; method and software development: GB, CFD, RS; code
 ⁵³⁴ review: GB, CFD, JF, RS; simulations and analyses: GB, CFD; writing original draft: GB and
 ⁵³⁵ CFD; reviewing and editing: GB, CFD, JF, RS, DPV.

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